

# Application of sugarcane bagasse for passive anaerobic biotreatment of sulphate rich wastewaters

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**Abstract** Biological treatment of sulphate-rich wastewaters employing dissimilatory sulphate reducing bacteria as remedial agents is an attractive technique and has gained importance in the last few years. Industrial effluents enriched with sulphates are generally deficient in electron donors. And thus cannot be treated biologically without supplementation of carbon through an external source. For scalable operations, however, the carbon source must not be expensive. In this context, present study reports the efficiency of biological sulphate reduction using sugarcane bagasse as a cost-effective carbon source. An average  $0.00391 \pm 0.001 \text{ gL}^{-1} \text{ day}^{-1}$  ( $3.91 \text{ mgL}^{-1} \text{ day}^{-1}$ ) sulphate reduction was observed reaching maximally to  $0.00466 \pm 0.001 \text{ gL}^{-1} \text{ day}^{-1}$  ( $4.66 \text{ mgL}^{-1} \text{ day}^{-1}$ ) while employing *Desulfovibrio fructosovorans*-HAQ2 and *Desulfovibrio piger*-HAQ6 in a 60-day trial of anaerobic incubation using sugarcane bagasse as growth substrate. These findings will be helpful in developing economical bioremediation processes tending to operate for a longer period of time to reduce sulphate contents of contaminated waters.

**Keywords** Carbon source · Economical bioremediation · Electron donor · Passive treatment · Sugarcane bagasse · Sulphate reduction

## Introduction

Several bioremediation processes including decontamination of acid mine drainage and sulphate rich waste waters

using sulphate-reducing bacteria (SRB) have been practiced much in the last decade (Johnson and Hallberg 2005; Neculita et al. 2007; Martins et al. 2009; Hussain and Qazi 2012). SRB collectively make a group of obligatory anaerobes exhibiting diverse morphological as well as physiological characteristics and occupy a wide range of habitats among terrestrial, sub-terrestrial and aquatic ecosystems (Willis et al. 1997). They may be autotrophic or heterotrophic with assimilatory or dissimilatory types of metabolisms. Both dissimilatory autotrophs and heterotrophs use sulphate ( $\text{SO}_4^{2-}$ ) as terminal electron acceptor. The former metabolize  $\text{CO}_2$  and the latter utilize multifarious energy-rich organic compounds as carbon sources/electron donors while generating sulfide (Pfennig et al. 1981; Lens and Kuennen 2001; Rabus et al. 2006; Martins et al. 2009). In various bioremediation processes based on the use of dissimilatory SRB, this biogenic sulfide reacts vigorously with dissolved metals present in the wastewaters forming insoluble precipitates of metal sulfides thus reducing the concentrations of metals and sulphates simultaneously (Costa and Duarte 2005; Vega-López et al. 2007; Martins et al. 2009).

Biological sulphate reduction is an energy-intensive process, thus, an efficient energy-rich carbon source is required (Barnes 1998). It is well known that SRB generally prefer simple low-molecular weight substrates such as sodium lactate and ethanol as energy sources. But being too much expensive, these cannot be used in bioremediation processes at large scales (Postgate 1984; Barnes 1998; El Bayoumy et al. 1999; Tsukamoto et al. 2004; Huisman et al. 2006). However, utilization of various environmental contaminants for instance, halogenated compounds and petroleum hydrocarbons' constituents has been reported by researchers (Fauque et al. 1991; Hao et al. 1996; Harms et al. 1999; Morasch et al. 2004). In addition, several

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different kinds of organic wastes have been employed in bioremediation processes using SRB and include animal manure, leaf mulch, molasses, mushroom compost, sawdust, sewage sludge, vegetal compost, watermelon rind, whey, wood chips and other agricultural wastes (Dvorak et al. 1992; Hammack et al. 1994; Christensen et al. 1996; Waybrant et al. 1998; Annachhatre and Suktrakoolvatt 2001; Costa and Duarte 2005; Coetser et al. 2006; Hussain and Qazi 2012; Hussain et al. 2014).

Selection of an economical and cost-effective growth substrate is of great significance in passive biotreatment processes. In most of the agricultural lands rice and wheat straw, rice husk and sugarcane bagasse are the key agricultural wastes. Of these wastes, sugarcane bagasse deserves special attention due to its abundant availability in almost all of the agricultural countries and competitive uses in many biotechnological processes especially as an ideal growth substrate for microbes for the production of value-added products such as bioethanol (Cardona et al. 2010), amino acids, enzymes, protein-rich animal feed and compounds of pharmaceutical importance (Parameswaran 2009). Application of sugarcane bagasse for various bioremediation processes is also well known (Krishnani et al. 2006; Garg et al. 2007; Sharma and Kaur 2011; Mukherjee et al. 2014). Sugarcane crop and its principal as well as byproducts are cheaper because the manual labour is far more cheaper and easily available whenever and wherever it is needed in developing countries, especially in Pakistan compared to developed and richer countries. Therefore, the cost-limiting factors such as collection, transportation and processing of sugarcane bagasse reinforce its utility in different biotechnological processes for such developing agricultural lands. Pakistan produces more than 12 million tons of the bagasse annually (Dawn News 2012). Currently, sugarcane bagasse is either used as fuel for electricity generation or contributes to environmental pollution in Pakistan (Ahmad and Qazi 2014). Utilization of sugarcane bagasse as growth substrate in biological sulphate reduction processes will lead to the concomitant treatment of both the organic and inorganic wastes. Keeping in view these facts, the present study was designed to investigate the efficiency of biological sulphate reduction using sugarcane bagasse as a cost effective and frequently available carbon source.

## Materials and methods

### Isolation of SRB from wastewater

Wastewater samples were collected from the bed of a leading domestic and industrial effluents carrying channel (Hadiara drain) in Lahore, Pakistan and processed for SRB culture enrichment as in Hussain and Qazi (2012) using

Postgate B medium (Postgate 1984). The composition of Postgate B medium is shown in Table 1. Sulfidogenic bacterial growth was assessed through the formation of black precipitates and production of rotten egg smell of  $H_2S$  which was checked by withdrawing and smelling the gas using a sterilized disposable syringe. The enrichments made thus were used to isolate pure cultures of SRB as described by Postgate (1984). Following the procedure, eight sulfidogenic bacterial strains were isolated and pure cultured.

### Batch experiments

These were performed in triplicates in serum bottles of 120 mL capacity using artificially prepared sulphate-rich

**Table 1** Compositions of various media used in this study

Medium	Ingredients	Quantity ( $g\text{L}^{-1}$ )
Postgate B medium	$\text{KH}_2\text{PO}_4$	0.5
	$\text{NH}_4\text{Cl}$	1
	$\text{CaSO}_4$	1
	$\text{MgSO}_4 \cdot 7\text{H}_2\text{O}$	2
	$\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$	0.5
	Sodium lactate	3.5 mL
	Yeast extract	1
	Ascorbic acid	0.1
	Thioglycollic acid	0.1 mL
	pH: 7.0–7.5	
Postgate E medium	$\text{KH}_2\text{PO}_4$	0.5
	$\text{NH}_4\text{Cl}$	1
	$\text{Na}_2\text{SO}_4$	1
	$\text{CaCl}_2 \cdot 6\text{H}_2\text{O}$	1
	$\text{MgCl}_2 \cdot 7\text{H}_2\text{O}$	2
	$\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$	0.5
	Sodium lactate	3.5 mL
	Yeast extract	1
	Ascorbic acid	0.1
	Thioglycollic acid	0.1 mL
Modified Postgate growth medium	Agar	15
	pH: 7.6	
	$\text{NH}_4\text{Cl}$	2
	$\text{CaCl}_2 \cdot 6\text{H}_2\text{O}$	0.06
	$\text{Na}_2\text{SO}_4$	2
	$\text{FeSO}_4 \cdot 7\text{H}_2\text{O}$	0.005
	Organic waste	2 %
	Yeast extract	0.1
	Sodium citrate	0.3
	pH: $7.0 \pm 0.5$	

wastewater which was actually a modification of Postgate growth medium and comprised of sulphate (2 g/L) and 2 % fine powder of sugarcane bagasse (w/v). Detailed composition of modified Postgate growth medium is shown in Table 1. Sodium lactate is the most widely used carbon source for cultivating SRB (Postgate 1984; Barnes 1998) and hence was employed in the control experiments. The inoculum size used was 5 % (v/v) harbouring around  $1.7 \times 10^6$  colony forming units (C.F.U.)/mL. pH of the medium was adjusted to 7.0 for each experiment. Diffusion of oxygen in inoculated media was prevented by adding a layer of autoclaved liquid paraffin (about 3–5 mm thick). The inoculated bottles were sealed with fine rubber stoppers and aluminium crimps and incubated at 30 °C for 60 days.

#### Periodic analysis of various parameters

Periodically (after every 10 days), 5 mL samples were withdrawn with the help of a sterilized syringe and filtered using a fine quality filter paper (Whatman, Cat No. 1001917, England). pH,  $\text{SO}_4^{2-}$  and C.F.U./mL were analyzed in each experiment. pH was measured with the help of a digital pH meter (InoLab, pH7110), while,  $\text{SO}_4^{2-}$  was estimated after Cha et al. (1999). The bacterial C.F.U. were estimated by cultivating 0.2 µL of a culture in Postgate E medium (Postgate 1984). The composition of this medium is provided in Table 1. The embedded black colonies were then counted after 5 days incubation at 30 °C.

#### Statistical analysis

Statistically the data were analysed using GLM procedures and mean values were compared using Duncan's Multiple Range test with the help of SAS 9.1. Differences between mean values will be considered significant at  $P < 0.05$ .

#### Genotypic characterization of the selected bacterial isolates

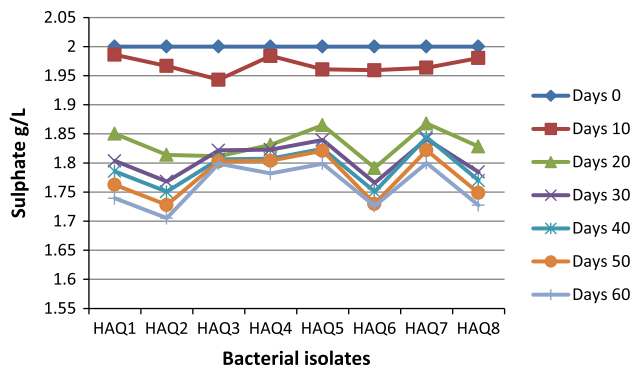
As present study reports sulphate reduction potentials of eight sulfidogenic bacterial isolates using sugarcane bagasse as growth substrate. In the initial screening based on sulphate reduction, the isolates which appeared a little bit different in terms of prodigious growth and better sulphate reduction performances were selected for genotypic characterization. For characterization of bacterial isolates at the molecular level, total genomic DNA was extracted from freshly grown cells of the selected bacterial isolates in Postgate B medium after Martins et al. (2009). Universal primers 27f (5'-AGAGTTTGATCMTGGCTCAG-3') and 1492r (5'-GGTTACCTTGTTACGACTT-3') were used to amplify 16S rRNA gene. PCR was performed in 50 µL

total reaction volume (DNA extract, 5 µL;  $\text{MgCl}_2$  (25 mM), 5 µL; dNTPs (1 mM), 5 µL; forward and reverse primers (5 pmol), 5 µL of each; DNA *Taq* polymerase, 2 U/mL; 1X *Taq* buffer, 5 µL; DNA free water, 18 µL). PCR amplification was carried out in a thermal cycler (Hamburg 22331, Germany) with a denaturation cycle for 3 min at 94 °C following 35 cycles of denaturation for 30 s at 95 °C, annealing step of 2 min at 60 °C and 1 min extension at 72 °C with a final extension step of 30 min at 72 °C. The PCR product obtained in this way was separated on 1 % agarose gel stained with ethidium bromide in TAE buffer by electrophoresis. Amplified bands of 1.5 kb were visualized under UV (Gel Doc, Bio-Rad Laboratories, USA) and excised for purification using Gene Purification Kit (Fermentas) following the manufacturer instructions. The amplicons were then got sequenced using Big Dye Terminator v3.1 cycle sequencing ready reactions (Macrogen, Korea) at the DNA Sequencing Facility, Korea. 16S rRNA gene sequences were assembled with phrap (version 0.990319). Homology searches were performed using BLAST (<http://www.ncbi.nlm.nih.gov/BLAST/>). The 16S rDNA sequences determined in this way were submitted to GenBank for obtaining accession numbers.

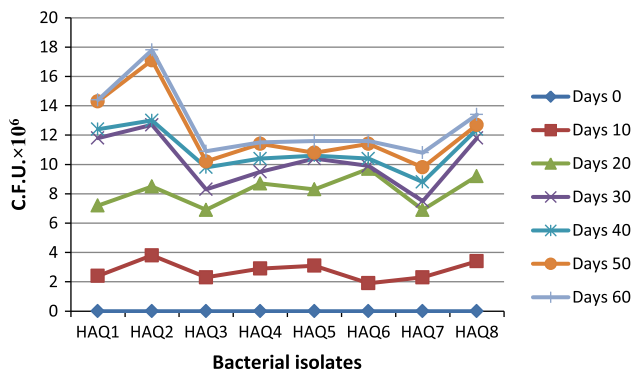
## Results and discussion

Genotypic characterization of the isolates, obtained in this study revealed that these isolates belonged to genus *Desulfovibrio*. Prevalence and dominance of *Desulfovibrio* genus in sulphate rich and anaerobic environments is best studied (Dar et al. 2005; Martins et al. 2009). Following the BLAST results the bacteria were identified as *Desulfovibrio fructosovorans* and *Desulfovibrio piger*. The GenBank has allotted accession numbers (KF536741 and KF536745, respectively) for these sequences. The growth patterns and sulphate reducing trends of all the isolates appeared very less variable throughout in this study.

Sulphates reduction rates remained comparatively higher in the first halves (between 10 to 30 days) and became gradually lower in the second halves (between 40 to 60 days) of observational periods (Fig. 1). Martins et al. (2009) have reported similar pattern of sulphate reduction during the assessment of food industry wastes as carbon sources in biological sulphate reduction processes. In spite of efficient sulphate reduction performances in the first halves, all the isolates showed extended lag phases showing negligible growth in the initial 10 days of anaerobic incubation periods (Fig. 2). SRB necessarily require longer lag periods to get acclimated to the more complex organic carbon sources (Beaulieu et al. 2000). Sugarcane bagasse is a heterogeneous mixture of simple and complex carbon polymers as presented in Table 2. Longevity of lag phases



**Fig. 1** Sulphate reduction trend shown by the SRB isolates using sugarcane bagasse as electron donor



**Fig. 2** Sulfidogenic bacterial growth pattern using sugarcane bagasse as carbon source

might be attributed to the acclimatization periods needed for SRB to utilize media containing complex carbon source following the induction and synthesis of proper enzyme system.

Once the bacterial growth became established in the presence of this complex substrate, sulphate reduction rates rose quickly to their peak values (between 10 to 20 days), dropped to their lower levels (from 20th day onwards) and gradually reduced to almost zero at the end of incubation periods (Fig. 1; Table 3). The earlier rise in sulphate reduction rates was probably due to consumption of frequently available simpler molecules in the primary stages of incubation. During the secondary stages of incubation, simple molecules might have been exhausted from the medium while promoting growth of SRB in the primary stages and, consequently, sulphate reduction rates decreased reasonably. The preference of SRB for simple organic molecules is a well reported phenomenon (Nagpal et al. 2000; Gibert et al. 2004; Tsukamoto et al. 2004; Zagury et al. 2006). When the SRB started to utilize this complex polymeric substrate, again increments in growth as well as sulphate reductions were recorded (Figs. 1, 2). For the complex nutrient media containing soluble as well as complex carbon sources, the process of diauxy is well

**Table 2** Physicochemical and elemental analysis of sugarcane bagasse

Content	Percent (w/w) on dry basis	Reference
Cellulose	40–60	McKendry (2002)
Hemicellulose	20–40	
Lignin	10–25	
Total sugars	13–18	Irfan et al. (2011)
Reducing sugars	8.02	
Crude fiber	45.61	Olagunju et al. (2014)
Crude proteins	3.3	
Lipid	2.01	
Volatile matter	84.83	Das et al. (2004)
Fixed carbon	13.28	
Ash	1.89	
C	56.32	
H	7.82	
N	0.89	
O	27.54	
Na	0.012	
K	0.175	
Ca	0.087	
Mg	0.437	
Al	0.003	
Fe	0.004	
Zn	0.001	
Cr	0.004	
Cu	0.001	
Mn	0.141	
P	0.014	
Si	0.911	
S	0.042	

C, H, N and O represent weight (%) on dry basis by difference

known (Crueger and Crueger 2005). These findings were consistent with those of Hussain and Qazi (2012) who reported similar results while studying the effectiveness of watermelon rind as carbon source in biological sulphate reduction.

In control experiments, 100 % sulphate reduction was observed when sodium lactate was used as electron donor. The total sulphate was reduced in the first 10 days of incubation. This efficient sulphate reduction was most probably due to simplicity of the lactate molecules and neutral pH of the media as have been reported earlier (Martins et al. 2009; Singh et al. 2011).

An average  $0.00391 \pm 0.001 \text{ gL}^{-1} \text{ day}^{-1}$  ( $3.91 \text{ mgL}^{-1} \text{ day}^{-1}$ ) sulphate reduction was observed reaching maximally to  $0.00466 \pm 0.001 \text{ gL}^{-1} \text{ day}^{-1}$  ( $4.66 \text{ mgL}^{-1} \text{ day}^{-1}$ ) while employing *Desulfovibrio fructosovorans*-HAQ2 and *Desulfovibrio piger*-HAQ6 using sugarcane bagasse as growth substrate (Table 4). An increased rate of sulphate reduction

**Table 3** Pattern of biological sulphate reduction following periodic incubation of 10 days using sugarcane bagasse as growth substrate

Bacterial isolate	Incubation period (days)						
	0	10	20	30	40	50	60
HAQ1	2.00 ± 0.00	1.99 ± 0.01 <sup>a</sup>	1.85 ± 0.02 <sup>bc</sup>	1.80 ± 0.01 <sup>cd</sup>	1.77 ± 0.01 <sup>cd</sup>	1.76 ± 0.01 <sup>c</sup>	1.75 ± 0.01 <sup>b</sup>
HAQ2	2.00 ± 0.00	1.97 ± 0.02 <sup>b</sup>	1.81 ± 0.00 <sup>e</sup>	1.77 ± 0.01 <sup>ef</sup>	1.75 ± 0.01 <sup>de</sup>	1.73 ± 0.01 <sup>d</sup>	1.72 ± 0.01 <sup>bc</sup>
HAQ3	2.00 ± 0.00	1.94 ± 0.01 <sup>c</sup>	1.83 ± 0.01 <sup>de</sup>	1.82 ± 0.01 <sup>bc</sup>	1.81 ± 0.01 <sup>b</sup>	1.80 ± 0.01 <sup>b</sup>	1.80 ± 0.01 <sup>a</sup>
HAQ4	2.00 ± 0.00	1.98 ± 0.02 <sup>a</sup>	1.84 ± 0.01 <sup>cd</sup>	1.82 ± 0.01 <sup>abc</sup>	1.81 ± 0.01 <sup>b</sup>	1.80 ± 0.01 <sup>b</sup>	1.78 ± 0.01 <sup>a</sup>
HAQ5	2.00 ± 0.00	1.96 ± 0.00 <sup>b</sup>	1.87 ± 0.00 <sup>ab</sup>	1.84 ± 0.00 <sup>ab</sup>	1.83 ± 0.01 <sup>ab</sup>	1.82 ± 0.01 <sup>a</sup>	1.80 ± 0.01 <sup>a</sup>
HAQ6	2.00 ± 0.00	1.96 ± 0.01 <sup>b</sup>	1.79 ± 0.00 <sup>f</sup>	1.77 ± 0.01 <sup>ef</sup>	1.75 ± 0.01 <sup>de</sup>	1.73 ± 0.01 <sup>d</sup>	1.72 ± 0.01 <sup>bc</sup>
HAQ7	2.00 ± 0.00	1.96 ± 0.01 <sup>b</sup>	1.87 ± 0.00 <sup>a</sup>	1.85 ± 0.01 <sup>a</sup>	1.84 ± 0.01 <sup>a</sup>	1.82 ± 0.01 <sup>a</sup>	1.80 ± 0.01 <sup>a</sup>
HAQ8	2.00 ± 0.00	1.98 ± 0.03 <sup>a</sup>	1.83 ± 0.01 <sup>de</sup>	1.78 ± 0.02 <sup>de</sup>	1.77 ± 0.01 <sup>cd</sup>	1.75 ± 0.01 <sup>c</sup>	1.75 ± 0.01 <sup>b</sup>

Values represent sulphate concentration (g/L) and are mean ± S.E. of three replicates. Those not sharing a common alphabet within a respective column are significantly different from each other

Single factor analysis of variance at  $P < 0.05$

**Table 4** Overall sulphate reduction rates of SRB isolates after completion of anaerobic incubation

Bacterial isolate	Total $\text{SO}_4^{2-}$ reduction (g/L <sup>-1</sup> )	$\text{SO}_4^{2-}$ reduction rate/day (g/L <sup>-1</sup> day <sup>-1</sup> )
HAQ1	0.25 ± 0.01	0.00416 ± 0.0001
HAQ2	0.28 ± 0.01	0.00466 ± 0.0001
HAQ3	0.20 ± 0.01	0.00333 ± 0.0001
HAQ4	0.22 ± 0.01	0.00366 ± 0.0001
HAQ5	0.20 ± 0.01	0.00333 ± 0.0001
HAQ6	0.28 ± 0.01	0.00466 ± 0.0001
HAQ7	0.20 ± 0.01	0.00333 ± 0.0001
HAQ8	0.25 ± 0.01	0.00416 ± 0.0001

Values are mean ± S.E. of three replicates

in the latter cases might be due to preference of this bacterial species particularly for consuming reducing sugars present in the growth substrate as reported earlier by Ollivier et al. (1988).

Sulphate reduction rates of all the isolates remained the same (0.0333 g/L<sup>-1</sup> day<sup>-1</sup>) as 100 % reductions were noticed in all the cases. The much lower reduction in case of sugarcane bagasse seemed significant when the cost of these two carbon sources was compared. Per kilogram costs of sodium lactate and sugarcane bagasse are 800 and 0.05 (in US Dollars), respectively.

Sulphate reduction rates of the pure cultures of SRB remained, in general, at low profiles while utilizing the lingo-cellulosic substrate, leaving behind a significant fraction of non-degraded carbon (cellulosic) mass which could be utilized by SRB if these were used in mixed cultures. An implication of mixed cultures is advantageous in providing bacterial consortium that facilitate both the development of reducing conditions as well as utilization of complex nutritive substrates (Gibert et al. 2002).

Sulphate reduction rates can be enhanced using bio-activated bacterial consortia. According to Beaulieu et al. (2000) bioactivation of bacterial consortia with an easily available simple organic source (e.g. lactate) and then replacing it with any organic waste or a mixture of organic wastes can lead to better sulphate reduction rates. It is also noticeable that multiple organic wastes perform better than a single waste (Waybrant et al. 1998; Zagury et al. 2006).

An important benefit of partially degradable carbon source is that it provides carbon for a long-term run and in this way fulfils the basic need of passive biotreatment processes (Reisman et al. 2003). Further work is required to delineate sulphate reduction potential of bioactivated and/or mixed SRB cultures employing multiple organic wastes including the one reported here.

## Conclusion

The availability, technical and economical viability of sugarcane bagasse and other similar lingo-cellulosic wastes (slowly biodegradable) in agricultural lands similar to Pakistan suggest these organic wastes as appropriate growth substrates for passive biotreatment of sulphate rich effluents/wastewaters than quite easily biodegradable but costly substrates. More better results can be obtained by diluting the wastewater to be treated. The efficiency of sulphate reduction can be enhanced further using mixed cultures of SRB as well as mixtures of different organic wastes. Thus, these aspects demand further studies in future.

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